

**Appendix G**  
Introduction to Noise

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# 1 Noise and Its Effect on People

Aircraft noise exposure in this document is primarily addressed using the Community Noise Equivalent Level (CNEL) metric. To assist reviewers in interpreting complex noise metrics, this appendix presents an introduction to the relevant fundamentals of acoustics and noise terminology, and the effects of noise on human activity.

## 1.1 Noise and its Metrics

Noise, often defined as unwanted sound, is one of the most common environmental issues associated with aircraft operations. Of course, aircraft are not the only sources of noise in an urban or suburban surrounding, where interstate and local roadway traffic, rail, industrial and neighbourhood sources may also intrude on the everyday quality of life. Nevertheless, aircraft are readily identifiable to those affected by their noise and are typically singled out for criticism. Consequently, aircraft noise problems often dominate analyses of environmental impacts.

A “metric” is defined as something “of, involving, or used in measurement.” As used in environmental noise analyses, a metric refers to the unit or quantity that quantitatively measures the effect of noise on the environment. Noise studies have typically involved a confusing proliferation of noise metrics used by individual researchers who have attempted to understand and represent the effects of noise. As a result, literature describing environmental noise or environmental noise abatement has included many different metrics.

Various federal agencies involved in environmental noise mitigation have agreed on common metrics for environmental impact analysis documents. Furthermore, the Federal Aviation Administration (FAA) has specified which metrics, such as CNEL and DNL, should be used for federal aviation noise assessments.

This section discusses the following acoustic terms and metrics:

- Decibel (dB)
- A-Weighted Decibel (dBA)
- Maximum Sound Level ( $L_{max}$ )
- Sound Exposure Level (SEL)
- Equivalent Sound Level ( $L_{eq}$ )
- Community Noise Equivalent Level (CNEL)
- Day-Night Average Sound Level (DNL)

### 1.1.1 The Decibel (dB)

All sounds come from a sound source—a musical instrument, a speaking voice, or an airplane passing overhead. It takes energy to produce sound. The sound energy produced by any sound source is transmitted through the air in sound waves—tiny, quick oscillations of pressure just

above and just below atmospheric pressure. These oscillations, or sound pressures, impinge on the ear creating the sound we hear.

Our ears are sensitive to a wide range of sound pressures. The loudest sound that we hear without pain has about one trillion times more energy than the quietest sounds we hear. On a linear scale, this range is unwieldy. Therefore, we compress the total range of sound pressures to a more meaningful range by introducing the concept of sound pressure level (SPL) and its logarithmic unit of decibel (dB).

SPL is a measure of the sound pressure of a given noise source relative to a standard reference value (typically the quietest sound that a young person with good hearing can detect). Decibels are logarithmic quantities—logarithms of the ratio of the two pressures, the numerator being the pressure of the sound source of interest, and the denominator being the reference pressure (the quietest sound we can hear).

The logarithmic conversion of sound pressure to SPL means that the quietest sound we can hear (the reference pressure) has a SPL of about zero decibels, while the loudest sounds we hear without pain have SPLs less than or equal to about 120 dB. Most sounds in our day-to-day environment have SPLs from 30 to 100 dB.

Because decibels are logarithmic quantities, they require logarithmic math and not simple (linear) addition and subtraction. For example, if two sound sources each produce 100 dB and are operated together, they produce only 103 dB—not 200 dB as might be expected. Four equal sources operating simultaneously result in a total SPL of 106 dB. In fact, for every doubling of the number of equal sources, the SPL (of all of the sources combined) increases another three decibels. A ten-fold increase in the number of sources makes the SPL increase by 10 dB. A hundredfold increase makes the level increase by 20 dB, and it takes a thousand equal sources to increase the level by 30 dB.

If one source is much louder than another, the two sources together will produce the same SPL (and sound to our ears) as if the louder source were operating alone. For example, a 100 dB source plus an 80 dB source produce 100 dB when operating together. The louder source “masks” the quieter one. But if the quieter source gets louder, it will have an increasing effect on the total SPL. When the two sources are equal, as described above, they produce a level 3 decibels above the sound level of either one by itself.

From these basic concepts, note that one hundred 80 dB sources will produce a combined level of 100 dB; if a single 100 dB source is added, the group will produce a total SPL of 103 dB. Clearly, the loudest source has the greatest effect on the total.

There are two useful rules of thumb to remember when comparing SPLs: (1) most of us perceive a 6 to 10 dB increase in the SPL to be an approximate doubling of loudness, and (2) changes in SPL of less than about 3 dB are not readily detectable outside of a laboratory environment.

### 1.1.2 A-Weighted Decibel (dBA)

Another important characteristic of sound is its frequency, or “pitch.” This is the rate of repetition of the sound pressure oscillations as they reach our ear. Frequency can be expressed in units of cycles per second (cps) or Hertz (Hz). Although cps and Hz are equivalent, Hz is the preferred scientific unit and terminology.

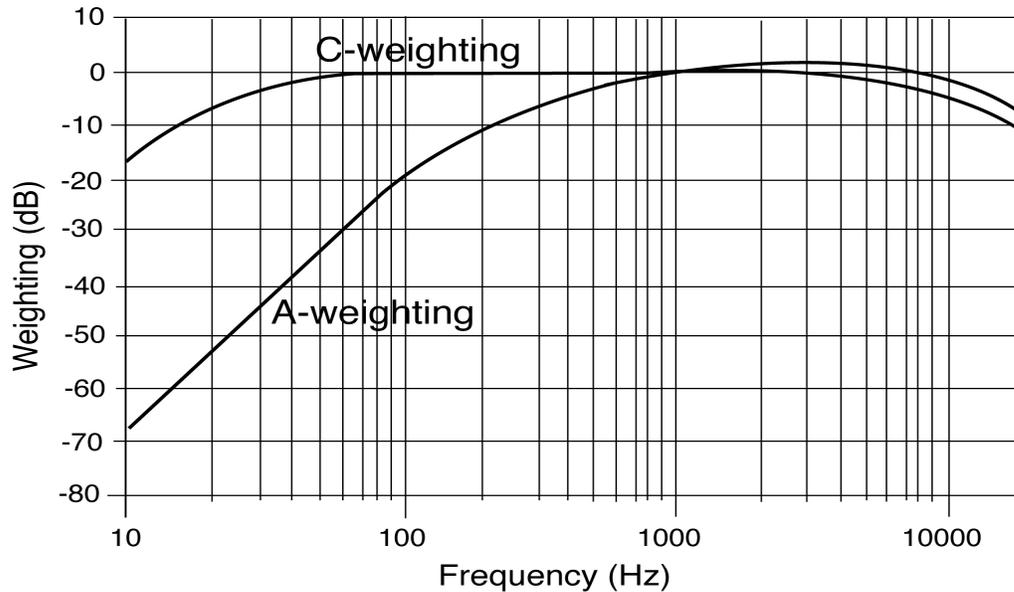
A very good ear can hear sounds with frequencies from 16 Hz to 20,000 Hz. However, most people hear from approximately 20 Hz to approximately 10,000-15,000 Hz. People respond to sound most readily when the predominant frequency is in the range of normal conversation, around 1,000 to 4,000 Hz. Acousticians have developed and applied “filters” or “weightings” to SPLs to match our ears’ sensitivity to the pitch of sounds and to help us judge the relative loudness of sounds made up of different frequencies. Two such filters, “A” and “C,” are most applicable to environmental noises.

A-weighting significantly de-emphasizes noise at low and high frequencies (below approximately 500 Hz and above approximately 10,000 Hz) where we do not hear as well. The filter has little or no effect at intervening frequencies where our hearing is most efficient. **Figure 1** shows a graph of the A-weighting as a function of frequency and its aforementioned characteristics. Because this filter generally matches our ears’ sensitivity, sounds having higher A-weighted sound levels are usually judged to be louder than those with lower A-weighted sound levels, a relationship which does not always hold true for unweighted levels. Therefore, A-weighted sound levels are normally used to evaluate environmental noise. SPLs measured through this filter are referred to as A-weighted decibels (dBA).

As shown in Figure 1, C-weighting is nearly flat throughout the audible frequency range, hardly de-emphasizing the low frequency noise. C-weighted levels are not used as frequently as A-weighted levels, but they may be preferable in evaluating sounds whose low-frequency components are responsible for secondary effects such as the shaking of a building, window rattle, perceptible vibrations or other factors that can cause annoyance and complaints. Uses include the evaluation of blasting noise, artillery fire, sonic boom, and in some cases, aircraft noise inside buildings. SPLs measured through this filter are referred to as C-weighted decibels (dBC).

Other weighting networks have been developed to correspond to the sensitivity and perception of other types of sounds, such as the “B” and “D” filters. However, A-weighting has been adopted as the basic measure of community environmental noise by the U.S. Environmental Protection Agency (EPA) and nearly every other agency concerned with aircraft noise throughout the United States.

Figure 1  
**Frequency Response Characteristics of Various Weighting Networks**



Source: ANSI S1.4-1983 "Specification of Sound Level Meters."

**Figure 2** presents typical A-weighted sound levels of several common environmental sources. Sound levels measured (or computed) using A-weighting are most properly called "A-weighted sound levels" while sound levels measured without any frequency weighting are most properly called "sound levels." However, since this document deals only with A-weighted sound levels, the adjective "A-weighted" will be hereafter omitted, with A-weighted sound levels referred to simply as sound levels. As long as the use of A-weighting is understood, there is no difference implied by the terms "sound level" and "A-weighted sound level" or by the dB or dBA units.

An additional dimension to environmental noise is that sound levels vary with time and typically have a limited duration, as shown in **Figure 3**. For example, the sound level increases as an aircraft approaches, then falls and blends into the background as the aircraft recedes into the distance (although even the background varies as birds chirp, the wind blows or a vehicle passes by). Sounds can be classified by their duration as continuous like a waterfall, impulsive like a firecracker or sonic boom or intermittent like an aircraft overflight or vehicle passby.

Figure 2  
Sound Levels of Typical Noise Sources (dBA)

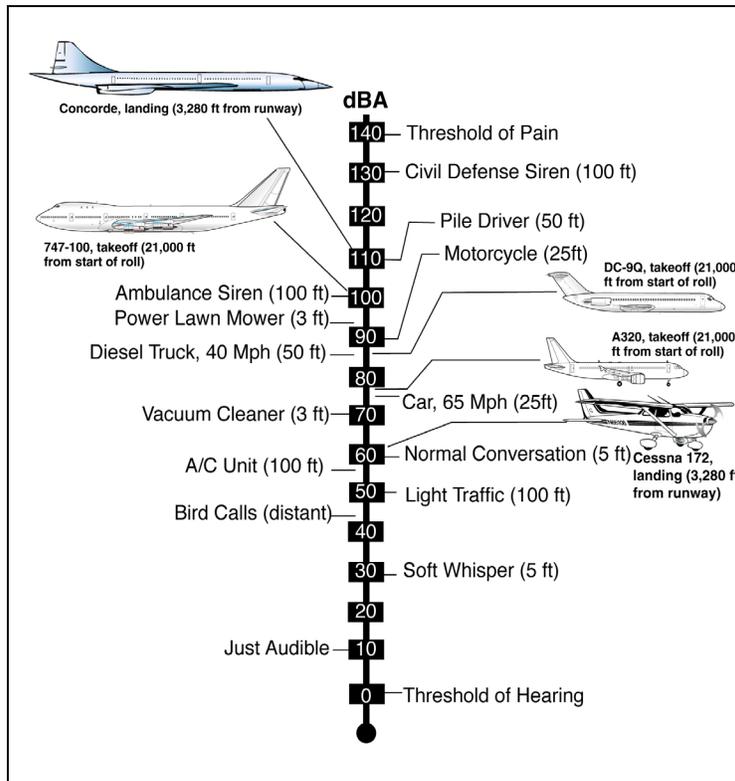
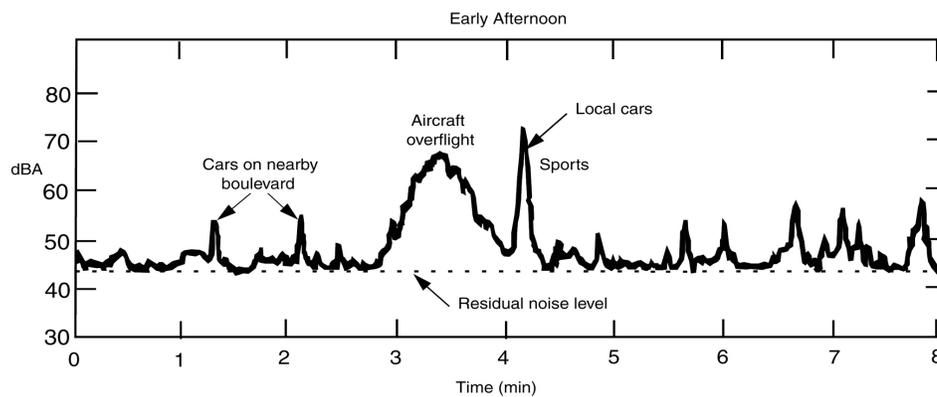


Figure 3  
Variation of Community Noise in a Suburban Neighborhood

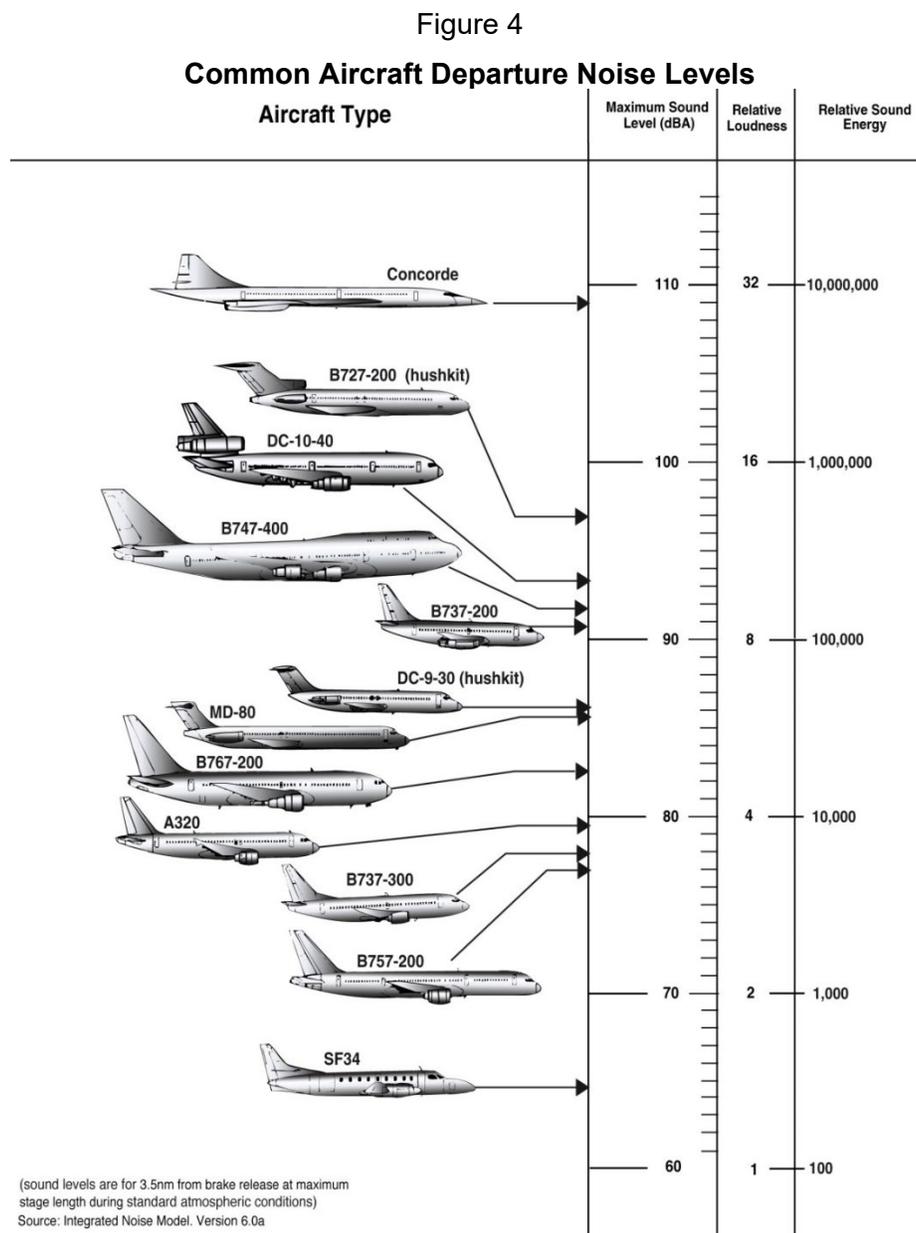


Source: "Community Noise," NTID 300.3 EPA, December 1971.

### 1.1.3 Maximum Sound Level ( $L_{max}$ )

The variation in sound level over time often makes it convenient to describe a particular noise “event” by its maximum sound level, abbreviated as  $L_{max}$ . For the aircraft overflight event in Figure 3, the  $L_{max}$  is approximately 67 dBA.

**Figure 4** shows  $L_{max}$  values for a variety of common aircraft from the FAA’s Integrated Noise Model (INM) database. These  $L_{max}$  values for each aircraft type are for aircraft performing a maximum stage (trip) length departure on a day with standard atmospheric conditions at a reference distance of 3.5 nautical miles (NM) from their brake release point. Of the dozen aircraft types listed on the figure, the Concorde has the highest  $L_{max}$  and the Saab 340 (SF340) has the lowest  $L_{max}$ .



The maximum level describes only one dimension of an event; it provides no information on the cumulative noise exposure generated by a sound source. In fact, two events with identical maxima may produce very different total exposures. One may be of short duration, while the other may continue for an extended period. The metric, discussed later in this appendix, corrects for this deficiency.

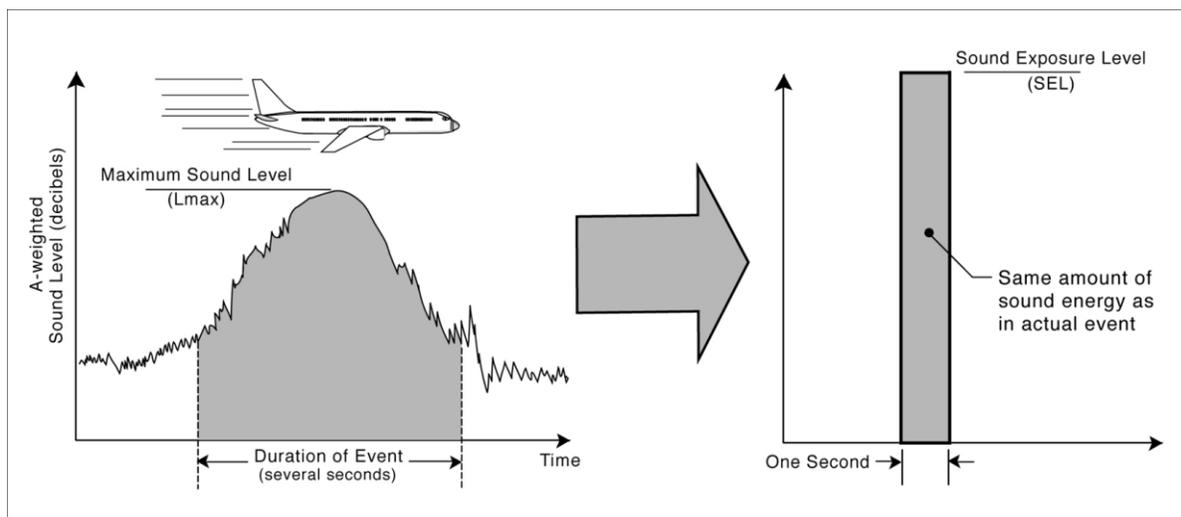
#### 1.1.4 Sound Exposure Level (SEL)

A frequently used metric of noise exposure for a single aircraft flyover is the Sound Exposure Level, or SEL. SEL may be considered an accumulation of the sound energy over the duration of an event. The shaded area in **Figure 5** illustrates that portion of the sound energy (or “dose”) included in an SEL computation. The dose is then normalized (standardized) to a duration of one second. This “revised” dose is the SEL, shown as the shaded rectangular area in Figure 5. Mathematically, the SEL represents the sound level of the constant sound that would, in one second, generate the same acoustic energy as the actual time-varying noise event. For events that last more than one second, SEL does not directly represent the sound level heard at any given time, but rather provides a measure of the net impact of the entire acoustic event.

Note that, because the SEL is normalized to one second, it will always be larger in magnitude than the maximum A-weighted level for an event that lasts longer than one second. In fact, for most aircraft overflights, the SEL is on the order of 7 to 12 dBA higher than the  $L_{max}$ . The fact that it is a cumulative measure means that not only do louder flyovers have higher SELs than quieter ones (of the same duration), but longer flyovers also have greater SELs than shorter ones (of the same  $L_{max}$ ).

It is the SEL’s inclusion of both the intensity and duration of a sound source that makes SEL the metric of choice for comparing the single-event levels of varying duration and maximum sound level. This metric provides a comprehensive basis for modeling a noise event in determining overall noise exposure.

Figure 5  
**Relationship Between Single Event Noise Metrics**



### 1.1.5 Equivalent Sound Level ( $L_{eq}$ )

Maximum A-weighted level and SEL are used to measure the noise associated with individual events. The following metrics apply to longer-term cumulative noise exposure that often includes many events.

The first cumulative noise metric, the Equivalent Sound Level (abbreviated  $L_{eq}$ ), is a measure of the exposure resulting from the accumulation of A-weighted sound levels over a particular period of interest (e.g., an hour, an 8-hour school day, nighttime or a full 24-hour day). However, because the length of the period can be different depending on the time frame of interest, the applicable period should always be identified or clearly understood when discussing the metric. Such durations are often identified through a subscript, for example  $L_{eq(8)}$  or  $L_{eq(24)}$ .

As for its application to aircraft noise issues,  $L_{eq}$  is often presented for consecutive 1-hour periods to illustrate how the hourly noise dose rises and falls throughout a 24-hour period, as well as how certain hours are significantly affected by a few loud aircraft. Since the period of interest for this study is in a full 24-hour day,  $L_{eq(24)}$  is the proper nomenclature.

Conceptually,  $L_{eq}$  may be thought of as a constant sound level over the period of interest that contains as much sound energy as the actual time-varying sound level with its normal “peaks” and “valleys,” as illustrated in Figure 3. In the context of noise from typical aircraft flight events and as noted earlier for SEL,  $L_{eq}$  does not represent the sound level heard at any particular time, but rather represents the total sound exposure for the period of interest. Also, it should be noted that the “average” sound level suggested by  $L_{eq}$  is not an arithmetic value, but a logarithmic, or “energy-averaged,” sound level. Thus, loud events tend to dominate the noise environment described by the  $L_{eq}$  metric.

### 1.1.6 Community Noise Equivalent Level (CNEL)

The Community Noise Equivalent Level (CNEL) metric is the predominant noise assessment metrics used in California. CNEL is the same as  $L_{eq}$  (an energy-average noise level over a 24-hour period) except that:

- 5 dB is added to those noise events occurring at evening (between 7 p.m. and 10 p.m.).
- 10 dB is added to those noise events occurring at night (between 10 p.m. and 7 a.m.).

This weighting reflects the added intrusiveness of evening and nighttime noise events attributable to the fact that community background noise levels decrease during these hours. CNEL does not represent the sound level heard at any particular time, but rather represents the total (and partially weighted) sound exposure.

### 1.1.7 Day-Night Average Sound Level (DNL)

DNL is the predominant noise assessment metrics used in states except California. DNL is similar with CNEL with the exception of 5 dB penalty applied to evening hours. Typical DNL values for a variety of noise environments are shown in **Figure 6** to indicate the range of noise exposure levels usually encountered.

Due to the DNL metric's excellent correlation with the degree of community annoyance from aircraft noise, DNL has been formally adopted by most federal agencies for measuring and evaluating aircraft noise for land use planning and noise impact assessment. Federal interagency committees such as the Federal Interagency Committee on Urban Noise (FICUN) and the Federal Interagency Committee on Noise (FICON) which include the EPA, FAA, Department of Defense, Department of Housing and Urban Development (HUD), and Veterans Administration, found DNL to be the best metric for land use planning. They also found no new cumulative sound descriptors or metrics of sufficient scientific standing to substitute for DNL. Other cumulative metrics could be used only to supplement, not replace DNL. Furthermore, FAA Order 1050.1E for environmental documents requires that DNL be used in describing cumulative noise exposure and in identifying aircraft noise/land use compatibility issues.<sup>1 2 3 4 5</sup>

Measurements of DNL are practical only for obtaining values for a relatively limited number of points. Instead, many noise studies, including this document, are based on estimates of DNL using an FAA-approved computer-based noise model.

Figure 6

### Typical Day-Night Average Sound Levels



Source: Federal Agency Review of Selected Airport Noise Analysis Issues, Federal Interagency Committee on Noise, August 1992.

## 1.2 The Effects of Aircraft Noise on People

To many people, aircraft noise can be an annoyance and a nuisance. It can interfere with conversation and listening to television, disrupt classroom activities in schools and disrupt sleep. Relating these effects to specific noise metrics aids in the understanding of how and why people react to their environment. This section addresses three ways we are potentially affected by aircraft noise: annoyance, interference of speech and disturbance of sleep.

### 1.2.1 Community Annoyance

The primary potential effect of aircraft noise on exposed communities is one of annoyance. The U.S. EPA defines noise annoyance as any negative subjective reaction on the part of an individual or group.<sup>1</sup>

Scientific studies<sup>1 2 3 6 7</sup> and a large number of social/attitudinal surveys<sup>8 9</sup> have been conducted to appraise the U.S. and inter-national community of annoyance due to all types of environmental noise, especially aircraft events. These studies and surveys have found the DNL to be the best measure of that annoyance.

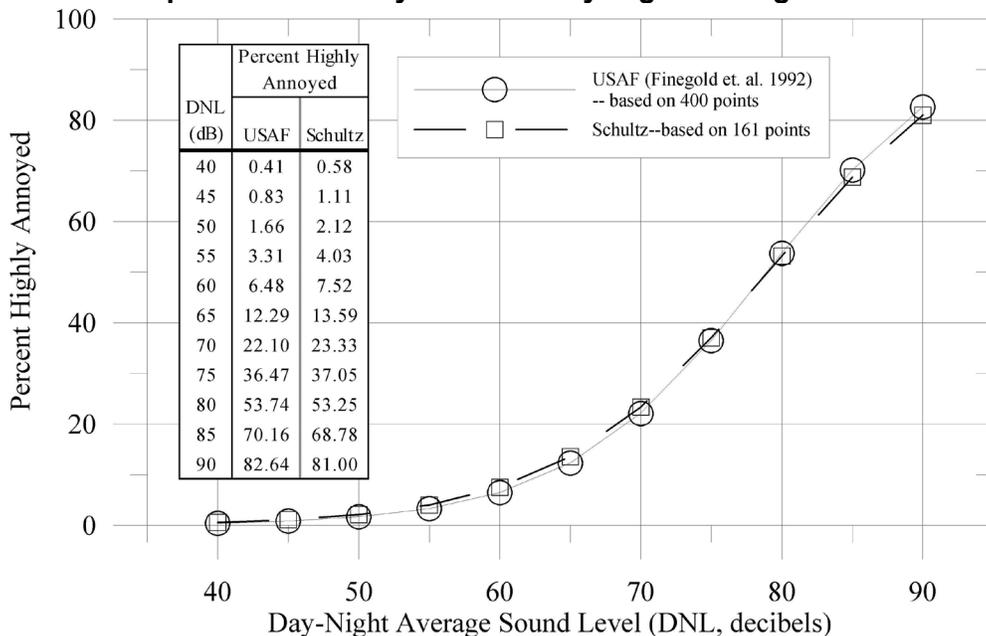
This relation between community annoyance and time-average sound level has been confirmed, even for infrequent aircraft noise events.<sup>10</sup> For helicopter overflights occurring at a rate of 1 to 52 per day, the stated reactions of community individuals correlated with the daily time-average sound levels of the helicopter overflights.

The relationship between annoyance and DNL that has been determined by the scientific community and endorsed by many federal agencies, including the FAA, is shown in **Figure 7**. Two lines in Figure 7 represent two large sets of social/ attitudinal surveys: one for a curve fit of

161 data points compiled by an individual researcher, Ted Schultz, in 1978<sup>8</sup> and one for a curve fit of 400 data points (which include Schultz's 161 points) compiled in 1992 by the U.S. Air Force.<sup>11</sup> The agreement of these two curves simply means that when one combines the more recent studies with the early landmark surveys in 1978, the results of the early surveys (i.e., the quantified effect of noise on annoyance) are confirmed.

Figure 7

Relationship Between Annoyance and Day-Night Average Sound Level



Source: Federal Interagency Committee on Noise (FICON),  
"Federal Agency Review of Selected Airport Noise Analysis Issues",  
August 1992, p. 3-6, Figure 3.1

**Figure 7** shows the percentage of people “highly annoyed” by a given DNL. For example, the two curves in the figure yield a value of about 13% for the percentage of people that would be highly annoyed by a DNL exposure of 65 dB. The figure also shows that at very low values of DNL, such as 45 dB or less, 1% or less of the exposed population would be highly annoyed. Furthermore, at very high values of DNL, such as 90 dB, more than 80% of the ex-posed population would be highly annoyed.

Recently, the use of DNL has been criticized as not accurately representing community annoyance and land-use compatibility with aircraft noise. One frequent criticism is based on the inherent feeling that people react more to single noise events and not as much to “meaningless” time-average sound levels. In fact, a time-average noise metric, such as DNL, takes into account both the noise levels of all individual events which occur during a 24-hour period and the number of times those events occur. As described briefly above, the logarithmic nature of the decibel unit causes the noise levels of the loudest events to control the 24-hour average.

As a simple example of this characteristic, consider a case in which only one aircraft overflight occurs in daytime hours during a 24-hour period, creating a sound level of 100 dB for 30 seconds. During the remaining 23 hours 59 minutes and 30 seconds of the day, the ambient sound level is 50 dB. The DNL for this 24-hour period is 65.5 dB. As a second example, assume that 10 such 30-second overflights occur in daytime hours during the next 24-hour period, with the same ambient sound level of 50 dB during the remaining 23 hours and 55 minutes of the day. The DNL for this 24-hour period is 75.4 dB. Clearly, the averaging of noise over a 24-hour period does not ignore the louder single events and tends to emphasize both the sound levels and number of those events. This is the basic concept of a time-average sound metric, and, specifically, the CNEL/DNL.

It is often suggested that a lower DNL, such as 60 or 55 dB, be adopted as the threshold of community noise annoyance for FAA environmental analysis documents. While there is no technical reason why a lower level cannot be measured or calculated for comparison purposes, a DNL of 65 dB:

- Provides a valid basis for comparing and assessing community noise effects.
- Represents a noise exposure level that is normally dominated by aircraft noise and not other community or nearby highway noise sources.
- Reflects the FAA's threshold for grant-in-aid funding of airport noise mitigation projects.
- HUD also established a DNL standard of 65 dB for eligibility for federally guaranteed home loans.

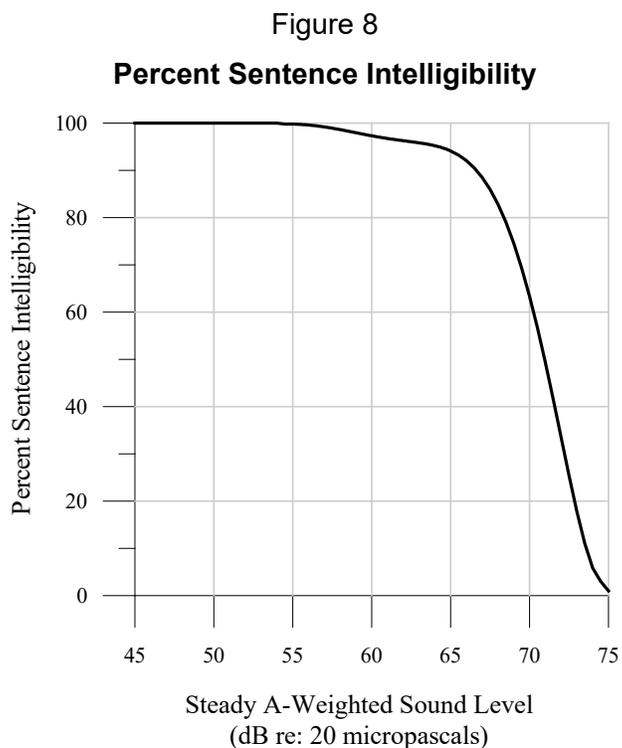
### 1.2.2 Speech Interference

A primary effect of aircraft noise is its tendency to drown out or “mask” speech, making it difficult to carry on a normal conversation.

Speech interference associated with aircraft noise is a primary cause of annoyance to individuals on the ground. The disruption of routine activities, such as radio or television listening, telephone use or family conversation, causes frustration and aggravation. Research has shown that “whenever intrusive noise exceeds approximately 60 dB indoors, there will be interference with speech communication.”<sup>1</sup>

Indoor speech interference can be expressed as a percentage of sentence intelligibility among two people speaking in relaxed conversation approximately one meter apart in a typical living room or bedroom.<sup>1</sup> The percentage of sentence intelligibility is a non-linear function of the (steady) indoor background sound level, as shown in **Figure 8**. This curve was digitized and curve-fitted for the purposes of this document. Such a curve-fit yields 100 percent sentence intelligibility for background levels below 57 dB and yields less than 10 percent intelligibility for background levels above 73 dB. Note that the function is especially sensitive to changes in sound level between 65 dB and 75 dB. As an example of the sensitivity, a 1 dB increase in background sound level from 70 dB to 71 dB yields a 14 percent decrease in sentence intelligibility.

In the same document from which Figure 8 was taken, the EPA established an indoor criterion of 45 dB DNL as requisite to protect against speech interference indoors.



Source: EPA 1974

### 1.2.3 Sleep Disturbance

Sleep disturbance is another source of annoyance associated with aircraft noise. This is especially true because of the intermittent nature and content of aircraft noise, which is more disturbing than continuous noise of equal energy and neutral meaning.

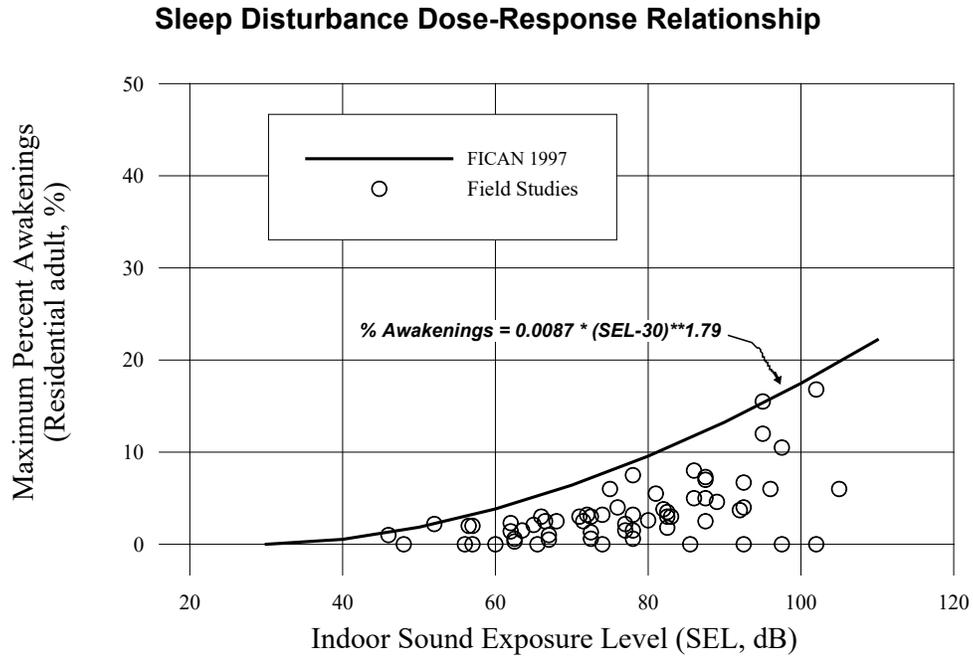
Sleep disturbance can be measured in one of two ways: “Arousal” represents awakening from sleep, while a change in “sleep stage” represents a shift from one of four sleep stages to another stage of lighter sleep without awakening. In general, arousal requires a higher noise level than does a change in sleep stage.

In terms of average daily noise levels, some guidance is available to judge sleep disturbance. The EPA identified an indoor DNL of 45 dB as necessary to protect against sleep interference.<sup>1</sup>

In June 1997, the Federal Interagency Committee on Aviation Noise (FICAN) reviewed the sleep disturbance issue and presented a sleep disturbance dose-response prediction curve.<sup>12</sup> FICAN based their curve on data from field studies<sup>13 14 15 16</sup> and recommends the curve as the tool for analysis of potential sleep disturbance for residential areas. **Figure 9** shows this curve which, for an indoor SEL of 60 dB, predicts that a maximum of approximately 5 percent of the residential

population exposed are expected to be behaviourally awakened. FICAN cautions that this curve should only be applied to long-term adult residents.

Figure 9



Source: FICA 1977

## **2 Airport Noise Modeling**

### **2.1 Introduction**

Noise levels in the vicinity of an airport can be modeled using the aircraft fleet, the time of day of operations, the runway orientation, layout, and utilization, representative noise model flight tracks and their respective utilization, aircraft performance data, weather and terrain input data. For projects that require federal actions, the FAA mandates the use of Aviation Environmental Design Tool (AEDT) to conduct aviation noise modeling. The CNEL metrics (See Section 1.1.6) was used as the primary noise metrics for this study.

### **2.2 Noise Modeling Software**

In 2015, the FAA released the Aviation Environmental Design Tool version 2b (AEDT 2b), which replaces both the INM and the Emissions and Dispersion Modeling System (EDMS), used for air quality analysis. The FAA issued a policy statement effective May 29, 2015 that required the use of AEDT 2b for new projects. Since the release of the AEDT 2b, the FAA has published several service packs that fixed various bugs and expanded its modeling capabilities. On September 12th, 2016, the FAA released AEDT version 2c (AEDT 2c) that incorporates various additional upgrades. On March 29<sup>th</sup>, 2021, the FAA released AEDT version 3d (AEDT 3d) with additional improvements, which is the most current version when this report was written.

### **2.3 Noise Metrics – CNEL and DNL**

In California, the CNEL is the predominant noise evaluation metrics which considers noise impacts during evening hours in addition to the DNL. Details on CNEL are included in Section 1.1.6. The DNL is the noise metric adopted by the Federal government to assess cumulative (i.e., multiple aircraft events) noise in the vicinity of airports for states other than California. Details on DNL are included in Section 1.1.7. The FAA has adopted the CNEL noise metrics as the standard noise metric for California. Therefore, in this analysis, aircraft noise is reported in terms of CNEL.

### **2.4 Operations**

#### **2.4.1 Average Annual Day (AAD)**

AEDT uses the Average Annual Day (AAD) to represent the time and frequency of flights at the airport. AAD operations are representative of all aircraft operations that occur over the course of a year, averaged over 365 days.

#### **2.4.2 Stage Length**

Stage length is a noise modeling term used to refer to trip distance for an aircraft departure from origin to destination, and is a surrogate for aircraft weight. The trip distance influences the take-off weight (and therefore the thrust and performance) of the aircraft, as more fuel is required to fly longer distances and therefore adds weight to the aircraft.

#### **2.4.3 Day/Evening/Night Split**

As described in Section 1.1.7, for CNEL, a 10 dB penalty is added to operations occur during the nighttime hours (between 10 p.m. and 7 a.m.) and a 5 dB penalty is added to operations occur during the evening hours (between 7 p.m. and 10 p.m.).

## **2.5 Runway and Track Utilization**

Runway use is a primary factor in the determination of noise exposure as how much each runway and helipad is utilized may determine the overall shape of the noise contour.

To determine projected noise levels on the ground, it is necessary to determine not only the frequency of aircraft operations, but also the altitude and location in which they fly. Flight routes to and from an airport, which are modeled as tracks in AEDT, are generally a function of the geometry of the airport's runways and the surrounding airspace structure in the vicinity of the airfield.

## **2.6 Maintenance Engine Run-ups**

Engine run-ups can be modeled in AEDT, and depending on their frequency, may influence the size and location of noise exposure contours

## **2.7 Terrain**

Terrain data is used to account for effects that variations in terrain have on noise propagation.

## **2.8 Weather**

The noise model allows for the modeling of atmospheric conditions in the calculation of noise exposure, taking into consideration temperature and humidity. Temperature is an important factor in aircraft performance, as higher temperatures decrease the density of air, which increases aircraft takeoff distance and reduces climb performance. This generally results in increased noise propagation in hot temperatures, as compared to colder temperatures.

## Endnotes

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- <sup>1</sup> U.S. Environmental Protection Agency, "Information on Levels of Environmental Noise Requisite to Protect the Public Health and Welfare with an Adequate Margin of Safety," Report 550/9-74-004, March 1974.
- <sup>2</sup> "Guidelines for Considering Noise in Land Use Planning and Control," Federal Interagency Committee on Urban Noise (FICUN), June 1980.
- <sup>3</sup> "Federal Agency Review of Selected Airport Noise Analysis Issues," Federal Interagency Committee on Noise (FICON), August 1992.
- <sup>4</sup> 14 CFR Part 150, Airport Noise Compatibility Planning, Amendment 150-3, Updated April 2012.
- <sup>5</sup> FAA Order 1050.1E, Chg 1, Environmental Impacts: Policies and Procedures, Department of Transportation, Federal Aviation Administration, March 20, 2006.
- <sup>6</sup> "Sound Level Descriptors for Determination of Compatible Land Use," American National Standards Institute Standard ANSI S3.23-1980.
- <sup>7</sup> "Quantities and Procedures for Description and Measurement of Environmental Sound, Part I," American National Standards Institute Standard ANSI S21.9-1988.
- <sup>8</sup> Schultz, T.J., "Synthesis of Social Surveys on Noise Annoyance," *J. Acoust. Soc. Am.*, 64, 377-405, August 1978.
- <sup>9</sup> Fidell, S., Barger, D.S., Schultz, T.J., "Updating a Dosage-Effect Relationship for the Prevalence of Annoyance Due to General Transportation Noise." *J. Acoust. Soc. Am.*, 89, 221-233, January 1991.
- <sup>10</sup> "Community Reactions to Helicopter Noise: Results from an Experimental Study," *J. Acoust. Soc. Am.*, 479-492, August 1987.
- <sup>11</sup> Finegold, L.S., C.S. Harris, H.E. VonGierke., "Applied Acoustical Report: Criteria for Assessment of Noise Impacts on People." *J. Acoust. Soc. Am.*, June 1992.
- <sup>12</sup> Federal Interagency Committee on Aviation Noise (FICAN), "Effects of Aviation Noise on Awakenings from Sleep," June 1997.
- <sup>13</sup> Pearson, K.S., Barber, D.S., Tabachnick, B.G., "Analyses of the Predictability of Noise-Induced Sleep Disturbance," USAF Report HSD-TR-89-029, October 1989.
- <sup>14</sup> Ollerhead, J.B., Jones, C.J., Cadous, R.E., Woodley, A., Atkinson, B.J., Horne, J.A., Pankhurst, F., Reyner, L, Hume, K.I., Van, F., Watson, A., Diamond, I.D., Egger, P., Holmes, D., McKean, J., "Report of a Field Study of Aircraft Noise and Sleep Disturbance." London Department of Safety, Environment, and Engineering, 1992.
- <sup>15</sup> Fidell, S., Pearsons, K., Howe, R., Tabachnick, B., Silvati, L., Barber, D.S. "Noise-Induced Sleep Disturbance in Residential Settings," AL/OH-TR-1994-0131, Wright Patterson AFB, OH, Armstrong Laboratory, Occupational and Environmental Health Division, 1994.
- <sup>16</sup> Fidell, S., Howe, R., Tabachnick, B., Pearsons, K., Sneddon, M., "Noise-Induced Sleep Disturbance in Residences Near Two Civil Airports," Langley Research Center, 1995.